

Bioplastic Degradation Potential of Microorganisms Isolated from the Soil

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Abstract: Plastic disposal is among the utmost problems threatening the environment at present, as enormous quantities of synthetic plastics are non-degradable. A constant exploration for bioplastic degraders is needed due to increasing demands for sustainable waste treatments. Bioplastic (Polycaprolactone) degradation potential of microorganisms isolated from soil was assessed. Soil sample was collected from a plastic dump site in Trans-Amadi, Port Harcourt, Nigeria. Isolation of Polycaprolactone (PCL) degrading bacteria and fungi were carried out via enrichment procedure and clear zone technique. Morphological, biochemical and molecular characteristics were used for strain identification of isolates. The optimal conditions for PCL degradation by bacterial and fungal strains were evaluated by growing the isolates at different temperatures (25, 30, 35 and 40°C), medium pH (ranging from pH 4 to 8) and salinities (5, 10, 15 and 20 ppt.) for 15 days. PCL residual weight in the culture was measured and used for PCL degradation assessment. Three PCL degrading bacterial strains (Bacillus megaterium, Alcaligenes aquatilis and Shewanella haliotis) and one fungal strain (Filobasidium uniguttalatum) were isolated. PCL maximum degradation rate of 59%, 56% and 53% respectively for bacterial strains and 62% for fungal strain were observed at 30°C. PCL maximum degradation rate of 63% for Filobasidium uniguttalatum were observed at pH 5 while the maximum degradation rate of 60% for Bacillus megaterium, 54% for Alcaligenes aquatilis and 57% for Shewanella haliotis were observed at pH 7. PCL maximum degradation rate of 58% for Bacillus megaterium, 55% for Alcaligenes aquatilis, 52% for Shewanella haliotis and 62% for Filobasidium uniguttalatum were observed at salinity of 5 ppt. It is concluded that the potential of the isolated indigenous microbial strains to degrade PCL qualifies them for use as compost cultures in waste treatment of bioplastic.

Keywords: Bioplastic, Polycaprolactone, Biodegradation, Synthetic Plastics, Bacteria, Fungi

1. Introduction

Biodegradable plastics are receiving major consideration as alternative to synthetic plastics whose increase as waste in the environment adversely affects the ecological unit [1, 2]. Hence plastic producers constantly optimize biodegradable products appropriate for domestic, industrial and agricultural purposes [3, 4]. But, in Nigeria, environmental pollution by plastic waste is on the increase and yet plastic manufacturers are yet to start production of biodegradable products. Polycaprolactone (PCL) is a petroleum-based polymer that can be degraded by microbes [5]. Nevertheless, accomplishing degradation in natural soils remains vague. Waste volume reduction is achieved when suitable microorganisms are applied during disposal of bioplastics leading to increased space availability in landfills [6]. Hence, inoculation with required microorganisms is essential for bioplastic degradation. The bioplastic degraders should be sourced for to lessen the buildup of plastic waste in the environs. Such microorganisms may possibly be added for biodegradation. Alcaligenes faecalis enhanced and Paecilomyces lilacinus were revealed to degrade PCL [7, 8]. PCL degradation by thermophilic actinomycetes [9] and Brevundimonas sp. strain MRL- ANI from soil [10] have been reported. Numerous anaerobic and aerobic polymerdegrading microbes have been isolated from seawater, soil and activated sludge [5]. There is no report of PCL degrading microbes inhabiting the Niger Delta environment.

Isolation and selection of indigenous strains able to degrade bioplastic to meet the precise requirement in Nigeria is needed. In this study, therefore, we assessed the polycaprolactone (PCL) degradation potential of bacteria and fungi isolated from a plastic dumpsite situated at Trans Amadi, Port Harcourt, Nigeria.

2. Materials and Methods

2.1. Sample Collection

Soil sample was collected from a plastic dumpsite at Trans Amadi Industrial Layout, Port Harcourt, Nigeria with coordinates N 4.817146°E 7.056100 (Figure 1).



Figure 1. Sample collection site.

2.2. Source of Bioplastic

Bioplastic (Poly-caprolactone pellets) was purchased from Sigma Aldrich USA, with average molecular weight 45,000, melting temperature of 60°C and Pcode 1002537759.

2.3. Enrichment and Isolation of Polycaprolactone– Degrading Bacterial and Fungal Strains

The polycaprolactone (PCL)-degrading bacteria and fungi were isolated via enrichment culture method and clear zone technique. Amodified mineral salt medium [11] containing (g/L of distilled water): KH_2PO_4 0.2, K_2HPO_4 0.8, $MgSO_4$ 0.1, $(NH_4)_2SO_4$ 0.4, $CaCl_2$ 0.4, $MnSO_4$ 0.01 and $FeSO_4$ 0.02 was employed. The final pH was adjusted to 4.5 for fungi and 7.2 for bacteria. The media were autoclaved at 121°C for 15 min. Then, 10 g of soil sample per 100ml of medium augmented with 1g/L of PCL as the only source of carbon in a 250ml Erlenmeyer flask was incubated for 7 days at 30°C. The enrichment technique was done using diverse concentrations (1, 2 and 5 g/L) of

polycaprolactone in the media. The bacteria and fungi present in the enhancement cultures were isolated on plates of mineral salt agar (pH 7.2 and 4.5 for bacteria and fungi respectively) augmented with 5 g/L polycaprolactone via spread plate method.

Isolates with clear zones were selected and streaked again and again on plates of nutrient agar (for bacteria) and acidified potato dextrose agar (for fungi) until pure. Morphological and biochemical characteristics [12] as well as molecular method were used for identification of purified bacterial isolates. The fungal strain was identified based on morphological characteristics [13] and phylogenetic analysis [14].

2.4. Molecular Characterization of the Microbial Isolates

2.4.1. Bacterial Genomic DNA Extraction

Five millilitres of overnight broth culture of bacterial isolate in Luria Bertani (LB) was spun at 14000 rpm for 3 min. The cells were re-suspended in 500 μ l of normal saline and heated at 95°C for 20 min. The heated bacterial suspension was cooled on ice and spun for 3 min at 14000 rpm. The supernatant containing the DNA was moved to 1.5 ml micro centrifuge tube and kept at -20°C for other downstream reactions.

2.4.2. DNA Quantification

The extracted genomic DNA was quantified using the Nanodrop 1000 spectrophotometer. The software of the equipment was launched by double clicking on the Nanodrop icon. The equipment was initialized with 2 μ l of sterile distilled water and blanked using normal saline. Two microlitres of the extracted DNA was loaded onto the lower pedestal, the upper pedestal was brought down to contact the extracted DNA on the lower pedestal. The DNA concentration was measured by clicking on the "measure" button.

2.4.3. 16S rRNA Amplification

The 16S rRNA region of the rRNA genes of the isolates was amplified using the 27F: 5'-AGAGTTTGATCMTGGCTCAG-3' 1492R: 5'and CGGTTACCTTGTTACGACTT-3' primers on an ABI 9700 Applied Biosystems thermal cycler at a final volume of 50 microlitres for 35 cycles. The PCR mix included: the X2 Dream taq Master mix supplied by Inqaba, South Africa (taq polymerase, DNTPs, MgCl), the primers at a concentration of 0.4 M and the extracted DNA as template.

The PCR conditions were as follows: Initial denaturation, 95° C for 5 minutes; denaturation, 95° C for 30 seconds; annealing, 52° C for 30 seconds; extension, 72° C for 30 seconds for 35 cycles and final extension, 72° C for 5 minutes. The product was resolved on a 1% agarose gel at 120V for 15 minutes and visualized on a UV transilluminator.

2.4.4. Fungal genomic DNA extraction

Extraction was done using a ZR fungal/bacterial DNA mini prep extraction kit supplied by Inqaba South Africa. A

heavy growth of the pure culture of the fungal isolate was suspended in 200 µl of isotonic buffer into a ZR Bashing Bead lysis tubes, 750 µl of lysis solution was added to the tube. The tubes were secured in a bead beater fitted with a 2ml tube holder assembly and processed at maximum speed for 5 minutes. The ZR bashing bead lysis tubes were centrifuged at 10,000 g for 1 minute. Four hundred (400) microlitres of supernatant was transferred to a Zymo-Spin IV spin, filtered (orange top) in a collection tube and centrifuged at 7000 g for 1 minute. One thousand two hundred (1200) microlitres of fungal/bacterial DNA binding buffer was added to the filtrate in the collection tubes bringing the final volume to 1600 µl, 800µlwas then transferred to a Zymo-Spin IIC column in a collection tube and centrifuged at 10,000g for 1 minute, the flow through was discarded from the collection tube. The remaining volume was transferred to the same Zymo-spin and spun. Two hundred (200) microlitres of the DNA Pre-Wash buffer was added to the Zymo-spin IIC in a new collection tube and spun at 10,000 g for 1 minute followed by the addition of 500 µl of fungal/bacterial DNA Wash Buffer and centrifuged at 10,000 g for 1 minute. The Zymo-spin IIC column was transferred to a clean 1.5 µl centrifuge tube, 100 µl of DNA elution buffer was added to the column matrix and centrifuged at 10,000 g for 30 seconds to elute the DNA. The ultra-pure DNA was then stored at -20°C for other downstream reaction.

2.4.5. Internal Transcribed Spacer (ITS) Amplification

The regions of the rRNA genes of the fungal isolate were amplified using the ITS1F: 5'-5'-CTTGGTCATTTAGAGGAAGTAA-3' and ITS4: TCCTCCGCTTATTGATATGC-3, primers on an ABI 9700 Applied Biosystems thermal cycler at a final volume of 50 microlitres for 35 cycles. The PCR mix included: the X2 Dream taq Master mix supplied by Inqaba, South Africa (taq polymerase, DNTPs, MgCl), the primers at a concentration of 0.4M and the extracted DNA as template.

The PCR conditions were as follows: Initial denaturation, 95°C for 5 minutes; denaturation, 95°C for 30 seconds; annealing, 53°C for 30 seconds; extension, 72°C for 30 seconds for 35 cycles and final extension, 72°C for 5 minutes.

The product was resolved on a 1% agarose gel at 120V for 15 minutes and visualized on a UV transilluminator. Sequencing were conducted using Big-Dye Terminator kit on a 3510 ABI sequencer by Inqaba Biotechnological, Pretoria South Africa. The sequencing was done at final volume of 10 μ l, the components included 0.25 μ l BigDye® terminator v1.1/v3.1, 2.25 μ l of 5 × BigDye sequencing buffer, 10 μ M Primer PCR primer, and 2-10ng PCR template per 100bp. The sequencing conditions were as follows 32 cycles of 96°C for 10s, 55°C for 5 s and 60°C for 4 min.

2.5. Phylogenetic Analysis

Obtained sequences were edited using the bioinformatics algorithm Trace edit, similar sequences were downloaded

from the National Center for Biotechnology Information (NCBI) data base using BLASTN. These sequences were aligned using ClustalX. The evolutionary history was inferred using the Neighbor-Joining method in MEGA 6.0 [15]. The bootstrap consensus tree inferred from 500 replicates [16] is taken to represent the evolutionary history of the taxa analyzed. The evolutionary distances were computed using the Jukes-Cantor method [17].

2.6. Effect of Temperature on Polycaprolactone Degradation by Bacterial and Fungal Strains

optimum temperature for polycaprolactone The degradation by bacterial and fungal strains was evaluated by growing the isolates at different temperatures from 25°C to 40°C. The isolates were cultured in 100 ml sterile mineral salt medium augmented with 5 g/L of PCL as the only source of carbon in a 250 ml Erlenmeyer flask. The test for each isolate was done in triplicate and 0.1 ml of a 24hr culture of the isolate was added to each flask. The uninoculated control and the flasks were incubated for 15 days at different temperatures (25, 30, 35, and 40°C). Degradation of PCL was evaluated by measuring residual weight of polycaprolactone in the culture. PCL percentage degradation was calculated via the equation [Wi-Wx/Wi] \times 100 where Wi is weight of PCL (mg) in the uninoculated control medium and Wx is the weight of PCL (mg) in the inoculated medium.

2.7. Effect of Initial Medium pH on PCL Degradation by Bacterial and Fungal Strains

The best medium pH for polycaprolactone degradation by bacterial and fungal strains was evaluated by cultivating the isolates at different medium pH from pH 4 to 8. The isolates were cultured in 100ml sterile mineral salt medium augmented with 5 g/L PCL as the only source of carbon in a 250ml Erlenmeyer flask. The test for each isolate was carried out in triplicate and 0.1ml of a 24hr culture of the isolate was added to the flask. The uninoculated control and the flasks were incubated for 15 days at 30°C. PCL residual weight in the culture was measured and used for PCL degradation evaluation.

2.8. Effect of Salinity on Polycaprolactone Degradation by Bacterial and Fungal Strains

The influence of salinity on PCL degradation by bacterial and fungal strains was evaluated by growing the isolates at different salinities (5, 10, 15, 20 ppt.). The isolates were cultured in 100ml sterile mineral salt medium augmented with 5 g/L PCL as the only source of carbon in a 250ml Erlenmeyer flask. The test for each isolate was carried out in triplicate and 0.1ml of a 24hr culture of the isolate was added to each flask. The uninoculated control flask and the flasks were incubated for 15 days sat 30°C. PCL residual weight in the culture was measured and used for PCL degradation estimation.

2.9. Statistical Analysis

Standard deviations for each of the experimental results were calculated using Excel Spreadsheets, with Microsoft excel software. Differences between treatments were examined for significance by one-way ANOVA and P = 0.05 was considered to be statistically significant.

3. Results and Discussion

3.1. PCL- Degrading Microbial Strains

A total of three PCL- degrading bacterial strains and one PCL- degrading fungal strain identified as *Alcaligenes aquatilis* (Figure 1), *Shewanella haliotis* (Figure 2), *Bacillus megaterium* (Figure 3) and *Filobasidium uniguttalatum* (Figure 4) respectively were isolated from soil via enrichment procedure and clear zone technique. Some strains of the genera isolated in this study such as *Bacillus brevis* 93 [18], *Bacillus pumilus* I-A [19] and *Shewanella* sp. CT01 [20] have been implicated as bioplastic – degrading bacteria. *Fusarium moniliforme* Fmm [21], *Paecilomyces verrucossum* [22], *Aspergillus* sp. XH0501-a [23] and *Penicillium oxalicum* D SYD05-1 [23] have been reported as bioplasticdegrading fungi.

Fifty eight isolates from agricultural soils in northern Thailand were implicated as PCL- degraders [24]. A thermophilic bacterium (*Ralstonia* sp. strain MRL-TL) isolated from hot spring degraded poly (ε - caprolactone) [25]. Hoang *et al.* [26] stated that 12 actinomycetes strains out of 305actinomycetes isolates degraded 3 polyesters [poly 3hydroxybutyrate (PHB), polyethylene succinate (PES) and PCL]. They stated that this capability could be recognized as a result of the clear zone around the isolates as also reported in this study. This microbial degradation of bioplastics has been ascribed to enzyme biosynthesis [27]. These strains isolated in this study could have potential for use in the management of bioplastic wastes. Figure 5 shows the Phylogenetic tree showing species relatedness of isolate F1.



0.05

Figure 2. Phylogenetic tree showing species relatedness of isolate B1.



Figure 3. Phylogenetic tree showing species relatedness of isolate B2.



Figure 4. Phylogenetic tree showing species relatedness of isolate B3.



0.05

Figure 5. Phylogenetic tree showing species relatedness of isolate F1.

3.2. Optimal Temperature for Polycaprolactone Degradation

The effect of temperature on polycaprolactone degradation by the bacterial and fungal strains is presented in Figure 6 PCL maximum degradation rate of 59% for *Bacillus megaterium*, 56% for *Alcaligenes aquatilis* 53% for *Shewanella haliotis* and 62% for *Filobasidium uniguttalatum* were observed at 30°C at the end of the 15 days. There was no growth in uninoculated sterile control indicating that the detected growth in the inoculated media happened at the expense of the bioplastic. A reduction in the degradation efficiency was detected at temperatures below and above 30°C. This indicates that temperature plays active role in microbial metabolism and polycaprolactone degradation. In Taiwan, Chau *et al.* [28] reported a bacterium (thermophilic *Streptomyces thermoviolaceus* subsp. *thermoviolaceus* 76T-2) which they isolated from soil and found to degrade poly (epsilon-caprolactone) (PCL) within 6 hr. at 45°C. Penkhrue *et al.* [24] stated that *Amycolatopsis* sp. strain SCM_MK2-4 which exhibited the maximum enzyme activity for poly lactic

acid (PLA) and PCL was capable of degrading 36.7% of the PLA in a culture medium after 7 days of incubation at 30°C. *Brevibacillusagri* degraded 84% of PCL originally supplemented in the medium at the end of 30 days incubation at 50°C (Nguyen and Trinh 2010). A thermophilic actinomycetes isolated from hot spring in Taiwan and identified as *Microbispora rosea* sub sp. *taiwanensis* fully degrade PCL, PHB and PES in liquid cultures at 50°C in 6 days [26].



Figure 6. Effect of temperature on PCL degradation by bacterial and fungal isolates.

3.3. Optimal Medium pH for Polycaprolactone Degradation

The influence of medium pH on polycaprolactone degradation is shown in Figure 7 PCL degradation occurred at a wide range of pH (4-8). PCL maximum degradation rate of 63% for Filobasidium uniguttalatum was observed at pH 5 while the maximum degradation rate of 60% for Bacillus megaterium, 54% for Alcaligenes aquatilis and 57% for Shewanella haliotis were observed at pH 7. A decline in the degradation efficacy was observed for pH lower and higher than 7 for bacteria and pH 5 for the fungus. The results denote that initial pH of culture medium affects the rate of PCL degradation. In Thailand, four polyester- degrading bacterial strains S14, Actinomadura sp. (Actinomadura sp. TF1. Streptomyces sp. APL3 and Laceyella sp. TP4) which were isolated from compost soils utilized polylactic acid (PLA), polycaprolactone (PCL), poly (butylene succinate) (PBS) and poly butylene succinate-co-adi-pate (PBSA) as substrates and exhibited robust depolymerase activity at pH 6.0-8.0 and temperatures of 40°C-60°C [30]. A PCL degrader, Brevibacillus agri, propagated well at pH 6-7 [31].

3.4. Optimal Salinity for Polycaprolactone Degradation

The effect of salinity on polycaprolactone degradation is shown in Figure 8 PCL degradation occurred at a wide range of salinity (0-20 ppt.). PCL maximum degradation rate of 58% for Bacillus megaterium, 55% for Alcaligenes aquatilis, 52% for Shewanella haliotis and 62% for Filobasidium uniguttalatum were detected at salinity of 5 ppt. A reduction in the degradation efficiency was detected for salinity lower and higher than 5 ppt. This demonstration that NaCl concentration plays active role in microbial metabolism polycaprolactone and degradation. Pseudomonas pachastrellae which grew well at 1.2 M NaCl, was found to be the main bacterium that degraded PCL in coastal environs [32]. Brevibacillus agri, a PCL degrader, grew well at1- 3% NaCl [31]. These observed variations in optimum conditions for PCL degradation may be due to the dissimilarities in source, nature and species of PCL degrading microbes. The variance in bioplastic degradation rates exhibited by different isolates in this study is not unexpected as diverse ecological factors are known to select for dissimilar microbial communities [33].



Figure 7. Effect of medium pH on PCL degradation by bacterial and fungal isolates.



Figure 8. Effect of salinity on PCL degradation by bacterial and fungal isolates.

4. Conclusion and Recommendation

In this study, three new PCL degrading bacterial strains (*Bacillus megaterium, Alcaligenes aquatilis* and *Shewanella haliotis*) and one new fungal strain (*Filobasidium uniguttalatum*) were isolated from soil. Their degradation rates at the end of 15 days were above 50%. The optimum temperature and salinity for PCL degradation were 30°C and 5 ppt. respectively. The maximum PCL degradation rate was observed at medium pH of 7 for bacterial strains and pH 5 for the fungal strain. These microorganisms have the potential to be employed as valuable tools for enhanced management of PCL- associated ecological waste.

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